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particulate smoke. The nano-structured incense soot particles intermixed with organics (e.g. formaldehyde and quinones) could increase the oxidative capacity. When considering the worldwide prevalence of incense burning and resulting high respiratory exposures, the oxygenated organics identified in this study have significant human health implications, especially for susceptible populations.

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57	Keywords separated by ' - '	Carbonyl - Combustion - Incense - Joss sticks - Polar organic - Quinone
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# Characterisation of airborne particles and associated organic components produced from incense burning

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Jennifer Bell · John Wenger · Kelly Bérubé

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**Abstract** Airborne particles generated from the burning of incense have been characterized in order to gain an insight into the possible implications for human respiratory health. Physical characterization performed using field emission-scanning electron microscopy showed incense particulate smoke mainly consisted of soot particles with fine and ultrafine fractions in various aggregated forms. A range of organic compounds present in incense smoke have been identified using derivatisation reactions coupled with gas chromatography-mass spectrometry analysis. A total of 19 polar organic compounds were positively identified in the samples, including the biomass burning markers levoglucosan, mannosan and galactosan, as well as a number of aromatic acids and phenols. Formaldehyde was among 12 carbonyl compounds detected and predominantly associated with the gas phase, whereas six different quinones were also identified in the incense particulate smoke. The nanostructured incense soot particles intermixed with organics (e.g. formaldehyde and quinones) could increase the

oxidative capacity. When considering the worldwide prevalence of incense burning and resulting high respiratory exposures, the oxygenated organics identified in this study have significant human health implications, especially for susceptible populations.

**Keywords** Carbonyl · Combustion · Incense · Joss sticks · Polar organic · Quinone

## Introduction

Particulate matter (PM) generated during combustion, and occurring in fine and ultrafine size fractions, is a common component of ambient air. Epidemiological studies have linked PM to adverse human health effects, such as cardiopulmonary diseases [1]. The respirable PM (i.e. PM<sub>2.5</sub>; less than 2.5 μm in aerodynamic diameter) has been implicated in causing adverse pulmonary effects in humans, given its ability to interact with the distal lung (e.g. alveoli) environment following inhalation [2]. During burning, in addition to the PM emission, organic compounds are released as gases, and the combination or synergistic effects of the two different phases may increase the risk of pulmonary disease [1]. Recently, combustion-related indoor activities have been the focus of inhalation toxicological and exposure assessments due to people spending the majority of their time indoors; these include incense and biomass burning, all of which can contribute to poor indoor air quality and become a public health concern [3]. The health end-points associated with exposure to air pollution are considered to be driven by the production of reactive oxygen species (ROS) initially in respiratory environment, causing inflammatory reactions and concomitant lung injury, as well as down-stream

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61 systemic disease (e.g. cardiopulmonary morbidity and  
62 mortality) [4].

63 Incense is commonly used during religious services and  
64 often in significant quantities. For example, in Kao-Hsiung  
65 city (Taiwan) in 2003, the Taiwan Environmental Protection  
66 Agency (EPA) calculated that approximately 28.7 metric  
67 tonnes of incense were burnt in temples [5]. In addition, an  
68 indication of the worldwide market for the incense trade  
69 between Asian countries (e.g. Taiwan) and the USA has  
70 been estimated to cost 4.6 million US dollars [6]. When  
71 considering incense burning in enclosed environments, the  
72 use of incense worldwide poses a serious public health  
73 issue, as currently identified by the World Health Organi-  
74 sation (WHO) [7].

75 Studies on the composition of PM from organic matter  
76 combustion have shown that a significant fraction of the total  
77 particulate emissions could be attributed to organic constitu-  
78 ents [8]. Polar organic compounds, including acids, alkanols,  
79 aldehydes and sugar derivatives, have been identified in fine-  
80 sized PM from various sources [9]. These organics, typically,  
81 have been found to contribute to more than 50% of the  
82 identified compounds and could comprise up to 80% of the  
83 solvent extractable organic chemicals in PM [9]. Carbonyls,  
84 many of which have been listed as air toxins, are usually  
85 found at higher concentration indoors when compared to  
86 outdoors, and are mainly emitted from combustion and  
87 building materials [8]. Most studies on incense PM organic  
88 fractions have been qualitative with some quantitative  
89 measurements of polycyclic aromatic hydrocarbon (PAH)  
90 [10], and hence, there is a paucity of studies on incense  
91 combustion-derived polar fractions such as oxidized deriv-  
92 atives of aromatic compounds (i.e. quinones) [11].

93 A range of oxygenated organic compounds in whole  
94 incense smoke have been investigated in previous studies  
95 using two-dimensional gas chromatography (GC×GC) [12–  
96 14], to show that the organic content of an aerosol may  
97 constitute a significant proportion of both particulate and gas  
98 phases triggering differential health effects post-exposure  
99 [15]. Consequently, when conducting an air toxins study,  
100 denuder-filter sampling techniques can be employed for the

101 rapid and simultaneous capture of PM and gas phase  
102 oxygenated compounds (e.g. carbonyl) [16]. Following  
103 oxygenated compound collection, derivatisation has com-  
104 monly been used for robust GC analysis. For example, *O*-  
105 (2,3,4,5,6-pentafluorobenzyl)-hydroxylamine (PFBHA) has  
106 been used to react with carbonyls with the following  
107 advantages: (1) better chromatographic separation, (2) higher  
108 sensitivity and (3) unknown compound identification [16].

109 Combustion-derived oxygenated organics in both particle  
110 and gas phases are of concern because of their high reactivity  
111 in the atmosphere and human lung environments [17]. The  
112 level of knowledge about incense-derived combustion prod-  
113 ucts contrasts with our knowledge of other combustion-  
114 derived indoor particles, such as cigarette smoke, which have  
115 been intensely researched [18]. To achieve the objectives in  
116 this study, incense smoke was physicochemically character-  
117 ized by field emission-scanning electron microscopy (FE-  
118 SEM) and GC/mass spectrometry (MS) to assess its potential  
119 implications for respiratory human health.

120 **Materials and methods**

121 **Sample collection**

122 Incense in the form of joss sticks, which are commonly used  
123 for religious ceremonies and to perfume air, were selected for  
124 this study. The samples consisted of four brands manufactured  
125 in Taiwan and three brands manufactured in China (Table 1,  
126 A–G). This study grouped these incense sticks by their main  
127 fragrant woods, including agarwood (Types A and B),  
128 sandalwood (Type C and D) and unknown (no information  
129 on the stem materials; Types E–G). The colour of the incense  
130 itself ranged from yellow to black (Table 1); however, there  
131 was no information on the organic components that made up  
132 the incense. The lengths, shapes and weights of the joss  
133 sticks are also given in Table 1.

134 All the incense burning ( $n=3$ ) was performed in a  
135 polyethylene terephthalate tube ( $0.0032\text{ m}^3$ ; length, 0.41 m;  
136 diameter, 0.1 m) mimicking a typical indoor burning

t1.1 **Table 1** Characteristics of the seven brands of joss sticks ( $n=6$ ; mean±SD)

t1.2	Incense ( $n=6$ )	Length of whole stick (cm)	Length of combustion part (cm)	Weight of whole stick (g)	Colour	Main material	Manufacture
t1.3	A	39.48±0.26	28.63±0.20	1.66±0.13	Dark yellow	Agarwood	Taiwan
t1.4	B	39.38±0.15	28.38±0.16	1.68±0.12	Dark yellow	Agarwood	Taiwan
t1.5	C	39.48±0.18	28.20±0.11	1.83±0.17	Black	Sandalwood	Taiwan
t1.6	D	39.58±0.08	28.83±0.12	1.42±0.08	Dark yellow	Sandalwood	Taiwan
t1.7	E	33.33±1.50	25.10±4.54	1.39±0.16	Yellow	Unknown	China
t1.8	F	34.00±0.06	24.35±0.19	1.32±0.16	Yellow	Unknown	China
t1.9	G	33.23±0.30	23.83±0.29	0.97±0.17	Black	Unknown	China

137 environment at an average of 25 °C and 55% relative  
 138 humidity (Fig. 1). The incoming clean air was conditioned  
 139 and purified by an absorption dryer (Ecodry KA-MT; Parker,  
 140 Charlotte, NC, USA), and was directed into the combustion  
 141 tube at a constant flow rate of 12 Lmin<sup>-1</sup> to maintain the  
 142 natural burning rate and to reduce heat accumulation inside  
 143 the tube. The incense stick was ignited, and the resultant  
 144 smoke was then passed through a five-channel annular  
 145 denuder (University Research Glassware, Chapel Hill, NC,  
 146 USA) coated with XAD-4 resin, followed by a 47-mm  
 147 quartz filter at a consistent flow rate of 10 Lmin<sup>-1</sup> for the  
 148 entirety of the 30-min burn. The denuder was used to capture  
 149 the gaseous organics; the filter was for particle collection.  
 150 After sampling, the denuder and filter were extracted with  
 151 appropriate solvents and subject to derivatisation and GC/  
 152 MS analysis. All the chemicals used in this study were  
 153 obtained from Sigma Aldrich Chemical Co. at the highest  
 154 purity available and used without further purification.

155 Field emission-scanning electron microscopy

156 A PM<sub>10</sub> selective-inlet head (horizontal elutriator; C30  
 157 Classifier; Thermo, UK) was used to collect PM onto a 47-  
 158 mm diameter polycarbonate filter (0.67 µm pores; Millipore,  
 159 UK) with a consistent flow rate at 30 Lmin<sup>-1</sup> [19]. Source-  
 160 specific particles were collected directly for 5 min. FE-SEM  
 161 (Philips Electron Optics, Eindhoven, Netherlands) was  
 162 undertaken on the samples using the methods of Bérubé et  
 163 al. [19] using filters mounted onto aluminium SEM stubs  
 164 (13 mm; Agar, UK), then imaged by FE-SEM at an  
 165 accelerating voltage of 25 kV, spot size 3.

166 BSTFA derivatisation

167 Polar organic compounds containing carboxyl and hydroxyl  
 168 groups were converted to their trimethylsilyl (TMS) deriva-  
 169 tives using *N,O*-bis(trimethylsilyl)trifluoroacetamide (BSTFA)  
 170 following the method of Kourtchev et al. [20]. The contents

of denuder and a quarter of the filter were extracted three times  
 with 20 ml dichloromethane:methanol (80:20 v/v) under  
 ultrasonic agitation for 30 min. Following reduction to  
 approximately 1 ml by rotary evaporation, the samples were  
 transferred to 2-ml amber vials and blown until dry using a  
 pure nitrogen stream. The residues were derivatised by the  
 addition of 60 µl of BSTFA containing 1% trimethylchlor-  
 osilane (BSTFA+1% TMCS) and 30 µl pyridine at 80 °C for  
 1 h [21]. The derivative samples were cooled to room  
 temperature, filtered and analyzed by GC/MS.

PFBHA derivatisation

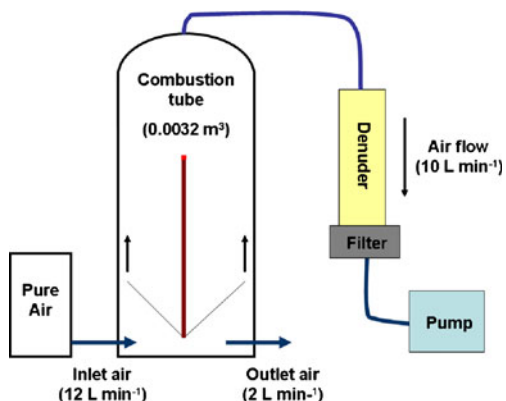
Carbonyls were converted to their oxime derivatives via  
 reaction with PFBHA [16, 22]. The contents of the denuder  
 and a quarter of the filter were extracted three times with  
 20 ml methanol and sonicated for 15 min. The samples  
 were subsequently derivatised with 10 mg PFBHA for 24 h  
 in the dark at room temperature. Of each sample, 0.1 ml  
 was dried under a nitrogen stream, and the residue was then  
 re-dissolved in 100 µl hexane in amber vials prior to  
 filtration for GC/MS analysis.

Derivatisation of quinones

Quinones were converted to their diacetyl derivatives by  
 reaction with acetic anhydride following the method of Cho et  
 al. [23]. The denuder and filter samples were sonicated in  
 dichloromethane for 15 min, reduced to 1 ml by rotary  
 evaporation and then reacted with 200 µl acetic anhydride  
 and 100 mg zinc as a catalyst at 80 °C for 15 min followed  
 by cooling to room temperature. The heating was repeated  
 for 15 min after adding a further 100 mg of zinc. The  
 reaction was again cooled to room temperature and  
 quenched by addition of 0.5 ml distilled water and 3 ml *n*-  
 pentane. Following centrifugation at 2,000 rpm for 10 min,  
 the *n*-pentane layer was filtered using PTFE membrane  
 filters (0.45 µm pore size) and transferred to clean amber  
 vials. The sample was evaporated until dry under a nitrogen  
 stream and then re-dissolved in 100 µl acetonitrile.

GC/MS analysis

All the samples were analysed after derivatisation using a GC/  
 MS instrument comprising a split/splitless injector (Varian  
 1079; Varian Inc., Palo Alto, CA, USA) interfaced to an ion  
 trap MS (Saturn 200; Varian Inc., Palo Alto, CA, USA). The  
 GC was equipped with a Rtx<sup>®</sup> 5MS fused silica capillary  
 column (Crossbond<sup>®</sup> 5% diphenyl, 95% dimethyl-  
 polysiloxane, 0.25 µm film thickness, 30 m in length and  
 0.25 mm i.d., Restek Corp., Bellefonte, PA, USA). Helium  
 was used as the carrier gas at a flow rate of 1 ml min<sup>-1</sup>. The  
 MS was operated at an electron energy of 70 eV and



**Fig. 1** Schematic diagram of the experimental set-up for collection of the gas and particle phase emissions from burning incense sticks

218 temperature of 200 °C using electron ionization (EI)  
 219 operated in full scan mode over the mass range 50–650 $m/z$   
 220 with a scan time of 1 s. Identification of the quinone  
 221 derivatives was also facilitated by operating the instrument  
 222 in single ion monitoring (SIM) mode ( $m/z$  110 for 1,4-  
 223 benzoquinone,  $m/z$  124 for methyl-1,4-benzoquinone,  $m/z$   
 224 222 for 2,6-di-tert-butyl-1,4-benzoquinone,  $m/z$  160 for 1,2-  
 225 naphthoquinone,  $m/z$  174 for methyl-1,4-naphthoquinone and  
 226  $m/z$  210 for 9,10-anthraquinone). The sample injection  
 227 volume was 1  $\mu$ l. The GC oven temperature programme  
 228 was identical to that reported in the literature for the different  
 229 groups of derivatives [16, 20, 22, 23]. Denuder and filter  
 230 blanks were prepared and analysed in the same manner to  
 231 determine the possible presence of contaminants. The  
 232 majority of compounds present in the denuder and filter  
 233 samples were identified by comparison with chromatograph-  
 234 ic retention times and mass spectra of authentic standards,  
 235 which were also prepared using the derivatisation procedures

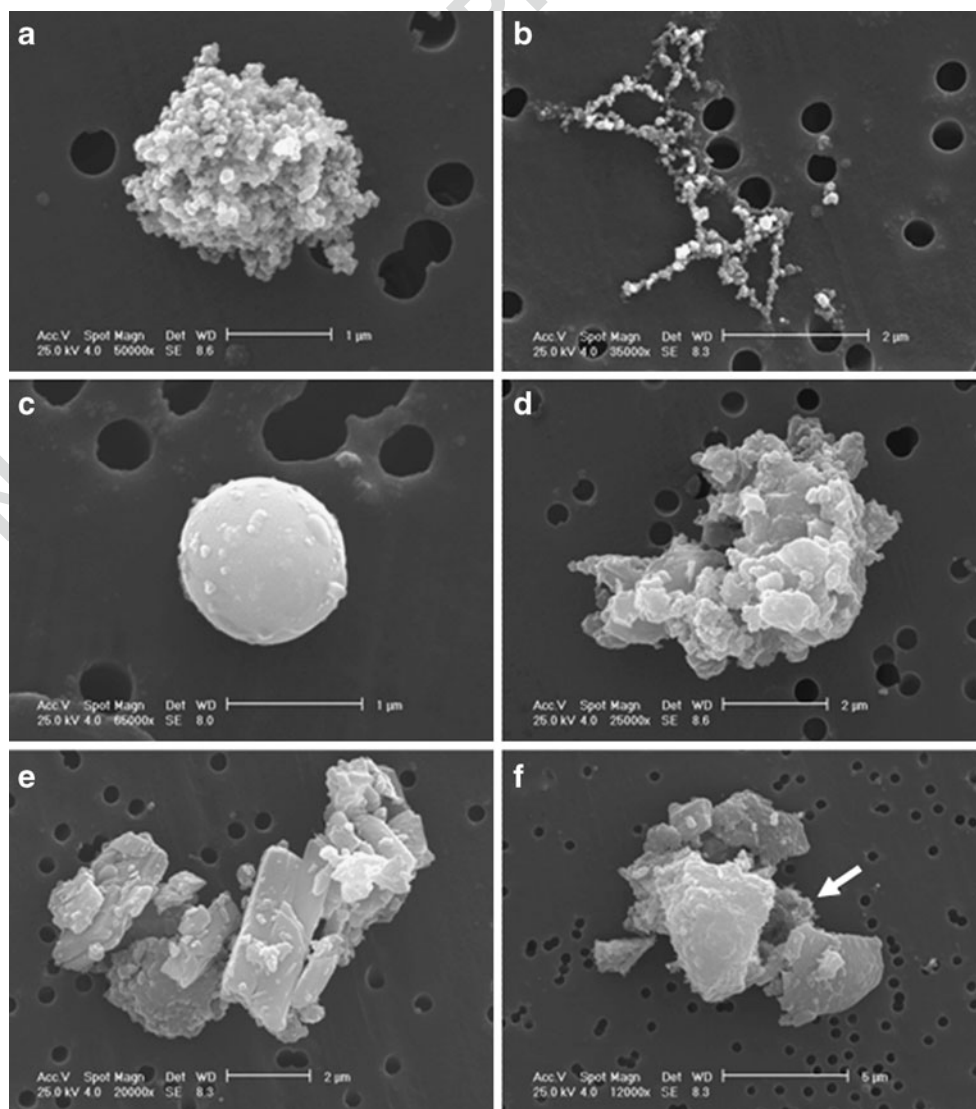
described above. Some compounds were also identified by  
 comparisons with a mass spectral library (National Institute  
 of Standards and Technology 05 mass spectra database).

## Results and discussion

### FE-SEM

PM generated by incense burning poses an important risk  
 for disease development due to its physicochemistry,  
 including its morphology that can be observed under  
 electron microscopy. The FE-SEM images show that soot  
 particles dominated the fine and ultrafine fractions, with the  
 individual soot nanospheres aggregated to form clusters  
 (Fig. 2a) or chains (Fig. 2b). A smaller number of larger  
 spherical soot-like particles were also observed in micro-sized  
 fractions (Fig. 2c). The second most abundant particle types

**Fig. 2** FE-SEM images of particles produced from incense burning. **a** Soot aggregate, developing from a chain into cluster. **b** Chain soot particles. **c** Larger spherical soot-like particle. **d** Irregular-shaped particle (clay). **e** Irregular-shaped particle aggregates. **f** Irregular-shaped particle aggregates and soot aggregates (arrow)





250 were irregular-shaped PM clusters with a ‘platy’ morphol- 275  
 251 ogy (Fig. 2d) or more ‘blocky’ shapes (Fig. 2e). These may 276  
 252 be attributed to unburnt incense components such as 277  
 253 organic materials (e.g. powdered charcoal) or inorganic 278  
 254 ‘filler’ minerals (e.g. finely powdered limestone). Compos-  
 255 ite particles were also commonly identified, often with  
 256 irregular-shaped particles acting as substrates for the soot  
 257 aggregates, which adhered to their surfaces (Fig. 2f, arrow).

258 Human respiratory disease is driven by the culmination 280  
 259 of acute and chronic exposure to intra- and extra-cellular 281  
 260 ROS generation [24]. There are various extra-cellular ROS 282  
 261 sources that can provoke directly or indirectly oxidative 283  
 262 stress. Many individual air pollutants that comprise an 284  
 263 ambient mixture are free radicals or have the ability to drive 285  
 264 free radical reactions. For instance, the oxidative stress 286  
 265 caused by PM and organics may arise from mixed sources, 287  
 266 involving (1) direct generation of ROS from the surface of 288  
 267 particles or the particle itself, (2) soluble compounds (e.g. 289  
 268 organic and inorganic compounds), (3) altered function of 290  
 269 mitochondria or NADPH-oxidase and (4) activation of 291  
 270 inflammatory cells capable of generating ROS and reactive 292  
 271 nitrogen species [24]. Of the identified particle types, soot 293  
 272 is of the greatest health concern because it can generate 294  
 273 large quantities of free radicals as a function of their small 295  
 274 size and large surface areas [25]. Biological ROS genera-

tion may also be correlated to PM type or morphology [26] 275  
 with spherical and irregular-shaped PM providing a 276  
 platform for inter-mixing or intercalating with organics that 277  
 cause redox cycling leading to apoptosis [24, 25]. 278

BSTFA derivatives 279

A total of 19 polar organic compounds (9 acids and 10 280  
 hydroxyls) was positively identified using BSTFA derivatisa- 281  
 tion (Table 2). The relative abundance of these compounds in 282  
 the gas and particle phases varied slightly from sample to 283  
 sample (Fig. 3). No consistency in the ratio of gas to PM (G/P) 284  
 abundance from incense burning was observed for either acids 285  
 or hydroxyls. In terms of the groups of incense-fragrant wood, 286  
 3-hydroxyprop-2-enoic acid, benzoic acid, 4-hydroxybutanoic 287  
 acid, vanillic acid, 1,4-cyclohexanediol, glycerine, 4- 288  
 hydroxybenzaldehyde, galactosan, mannosan and levoglucos- 289  
 an were determined in both particle and gas phases among 290  
 agarwood, sandalwood and the unknown group (Table 2). 291

Polar organics have been widely used in atmospheric 292  
 research for tracing and identifying specific sources in the 293  
 ambient air. Levoglucosan, mannosan and galactosan, for 294  
 example, are primary combustion products of cellulose and 295  
 hemicellulose and thus commonly used as tracers for 296  
 biomass burning [27]. Joss sticks are made of a wooden 297

t2.1 **Table 2** Polar organic compounds (acids and hydroxyls) identified in the gas (G) and particle (P) phase components of incense smoke (*n*=3)

Compound	G/P <sup>a</sup>	Incense types						
		Agarwood		Sandalwood		Unknown		
		G	P	G	P	G	P	
Acid								t2.2
6-Hydroxyhexanoic acid	1.4	V	–	V	V	V	V	t2.3
Propanoic acid	1.2	V	–	V	–	V	V	t2.4
3-Hydroxyprop-2-enoic acid	5.8	V	V	V	V	V	V	t2.5
Benzoic acid	7.2	V	V	V	V	V	V	t2.6
Glyceric acid	1.5	V	–	V	V	V	V	t2.7
4-Hydroxybutanoic acid	0.2	V	V	V	V	V	V	t2.8
Cinnamic acid	0.7	V	V	–	V	–	–	t2.9
3-Hydroxybenzoic acid	0.0	–	V	–	V	–	V	t2.10
Vanillic acid	0.2	V	V	V	V	V	V	t2.11
Hydroxyl								t2.12
Butanol	1.8	V	–	V	–	V	V	t2.13
1,4-Cyclohexanediol	0.4	V	V	V	V	V	V	t2.14
2,2'-Oxybisethanol	7.9	V		V	V	V	V	t2.15
Glycerin	1.1	V	V	V	V	V	V	t2.16
4-Hydroxybenzaldehyde	0.6	V	V	V	V	V	V	t2.17
<i>m</i> -Cresol	1.9	V	–	V	–	V	V	t2.18
Resorcinol	0.0	V	V	–	V		V	t2.19
Galactosan	0.1	V	V	V	V	V	V	t2.20
Mannosan	0.1	V	V	V	V	V	V	t2.21
Levoglucosan	0.2	V	V	V	V	V	V	t2.22

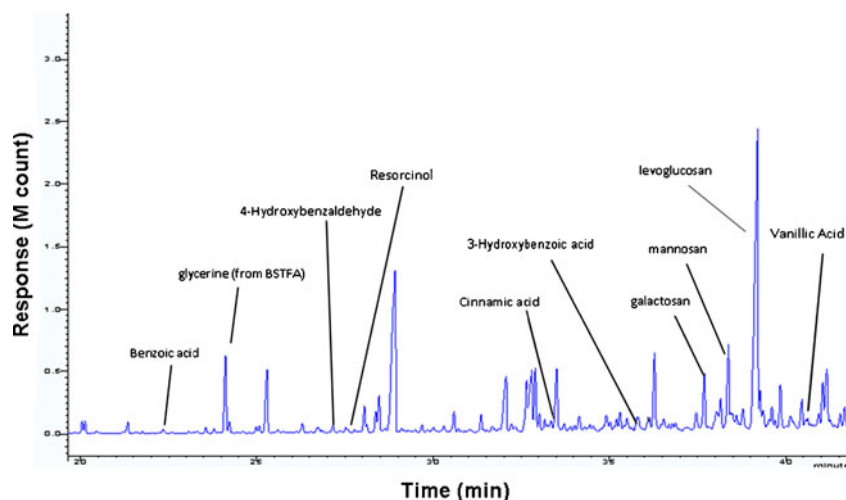
The incense has been classified into three types, agarwood, sandalwood and unknown, corresponding to the wood material used in the sticks

“–” compound not found in the sample, “V” compound found in the sample

<sup>a</sup> Relative abundance in the gas and particle phases

Q9

**Fig. 3** Total ion chromatogram showing BSTFA derivatives of particulate phase polar organic compounds (filter sample) produced from burning incense



298 stem coated with a range of incense materials including  
 299 aromatic wood, bark, herbs, seeds and flowers [28]. The  
 300 biomass-derived chemo-markers were detected as major  
 301 species in incense smoke [14]. The other polar organics  
 302 identified in this study are mainly oxygenated aromatic  
 303 compounds; a number of them have been observed in  
 304 incense smoke emissions and aromatic resins [12, 13, 29,  
 305 30]. Although it is difficult to assess the health implications  
 306 of exposure to the polar organics due to the paucity of  
 307 experimental information on the toxicology of these

compounds, some of these polar organic compounds could  
 potentially be used as chemo-markers of incense burning or  
 biomarkers following exposure [31].

PFBHA derivatives

The chromatograms of PFBHA derivatives obtained from  
 burning the different types of incense were very similar and  
 featured the same major peaks. There were 12 carbonyl  
 compounds identified in the samples (Table 3). The

t3.1 **Table 3** Carbonyls and quinones identified in the gas (G) and particle (P) phase components of incense smoke ( $n=3$ )

Compound	G/P <sup>a</sup>	Incense types						
		Agarwood		Sandalwood		Unknown		
		G	P	G	P	G	P	
Carbonyl								
Formaldehyde	10.4	V	V	V	V	V	V	t3.5
Acetaldehyde	5.4	V	V	-	V	-	V	t3.6
Acetone	9.7	V	V	V	V	V	V	t3.7
2-Butanone	3.4	V	V	V	V	V	V	t3.8
3-Methyl-2-butanone	5.8	V	V	V	V	V	V	t3.9
3-Pentanone	16.9	V	-	V	V	V	-	t3.10
2-Heptanone	-	V	-	V	-	V	-	t3.11
2-Butenal	-	V	-	V	-	V	-	t3.12
3(2H)-Pyridazinone	-	V	-	V	-	V	-	t3.13
Benzaldehyde	-	V	-	-	-	-	-	t3.14
Glyoxal	119.1	V	V	V	V	V	V	t3.15
Methylglyoxal	85.3	V	V	V	V	V	V	t3.16
Quinone								t3.17
1,4-BQ	0.7	V	V	V	V	V	V	t3.18
MBQ	0.1	V	V	V	V	V	V	t3.19
DTBQ	0.2	V	V	V	V	V	V	t3.20
1,2-NQ	0.2	V	V	-	-	-	-	t3.21
MNQ	0.1	V	V	V	V	V	V	t3.22
9,10-AQ	0.0	V	V	V	V	V	V	t3.23

The incense has been classified into three types, agarwood, sandalwood and unknown, corresponding to the wood material used in the sticks

“-” compound not found in the sample, “V” compound found in the sample

<sup>a</sup> Relative abundance in the gas and particle phases

316 intensity of the peaks in the denuder samples was  
 317 significantly larger than in the filter samples (3.4–119.1  
 318 times), indicating that the carbonyls were mainly present in  
 319 the gas phase (Table 3). For instance, some, such as 2-  
 320 heptanone, 2-butenal, 3(2H)-pyridazinone and benzalde-  
 321 hyde, were only found in the gas phase. However, several  
 322 carbonyls were observed in both phases from the three  
 323 groups of incense: formaldehyde, acetone, 2-butanone, 3-  
 324 methyl-2-butanone, glyoxal and methylglyoxal (Table 3).

325 A note in previous indoor research [8], incense burning is  
 326 an important source for human exposure to carbonyl  
 327 pollutants, which can cause poor indoor air quality and  
 328 increase the health concern of disease development. One of  
 329 the carbonyls in incense smoke, for example, has been firmly  
 330 recognized as a toxic substance by the US EPA: formalde-  
 331 hyde [32]. Formaldehyde is recognised as a level B1  
 332 probable human carcinogen, which is based on limited  
 333 evidence of carcinogenicity in humans. This chemical can  
 334 cause irritation to the skin, eyes and respiratory systems, and  
 335 reproductive diseases [32]. Furthermore, the high reactivity  
 336 leads carbonyls to generate free radicals in biological  
 337 environments and causes oxidative stress, DNA adducts  
 338 and inflammatory responses, such as formaldehyde [33].

339 Quinone derivatives

340 Six quinone standards were pre-selected for their important  
 341 properties and scanned for using GC/MS in the SIM mode:  
 342 1,4-benzoquinone (1,4-BQ), methyl-1,4-benzoquinone  
 343 (MBQ), 2,6-di-tert-butyl-1,4-benzoquinone (DTBQ), 1,2-  
 344 naphthoquinone (1,2-NQ), methyl-1,4-naphthoquinone  
 345 (MNQ) and 9,10-anthraquinone (9,10-AQ). Five of which  
 346 were found predominantly in the incense particle phase,  
 347 including 1,4-BQ, MBQ, 1,2-NQ and 9,10-AQ, as the most  
 348 abundant (Table 3).

349 Although quinones have previously been observed in  
 350 emissions arising from fuel combustion [23], this is the  
 351 first study in which these compounds have been identified  
 352 in incense smoke. Particulate quinones are highly reactive  
 353 and can bind covalently with nucleophilic regions of  
 354 proteins and DNA [31] and increase ROS formation by  
 355 redox cycling [34]. The macromolecule damage driven  
 356 by quinones initialled cell defence mechanism causing  
 357 apoptosis [34].

358 Our previous study determined that incense PM was able  
 359 to increase acellular ROS formation and oxidative DNA  
 360 damage [35], which could be driven by their organic  
 361 fractions (e.g. carbonyls and quinones) and their physical  
 362 characteristics (e.g. spherical soot). Therefore, particulate  
 363 quinones generated from incense burning are of particular  
 364 concern, as they may be able to penetrate into the distal  
 365 lung environments due to PM's respirable size and cause  
 366 adverse health effects.

**Conclusions**

367 A comprehensive characterisation of oxygenated organic  
 368 compounds generated by incense burning and their poten-  
 369 tial adverse health effects was undertaken. This study  
 370 identified a broad range of oxygenates, including polar  
 371 organics, carbonyls and quinones, in incense smoke that are  
 372 known critical indoor air pollutants. A number of the  
 373 compounds are known to be toxic, and information on their  
 374 relative abundance in the gas and particle phases has been  
 375 obtained. This study demonstrated that incense smoke may  
 376 pose a significant adverse risk to human health, and further  
 377 toxicology studies based on these organic compounds  
 378 should be the subject of extensive investigations.  
 379

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